1	Rapid multi-generational acclimation of coralline algal reproductive structures to ocean
2	acidification
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 variability

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31 Abstract

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33 The future of coral reef ecosystems is under threat because vital reef-accreting species such as 34 coralline algae are highly susceptible to ocean acidification. Although ocean acidification is known to 35 reduce coralline algal growth rates, its direct effects on the development of coralline algal 36 reproductive structures (conceptacles) is largely unknown. Furthermore, the long-term, multi-37 generational response of coralline algae to ocean acidification is extremely understudied. Here, we 38 investigate how mean pH, pH variability and the pH regime experienced in their natural habitat affect 39 coralline algal conceptacle abundance and size across six generations of exposure. We show that 40 second generation coralline algae exposed to ocean acidification treatments had conceptacle 41 abundances 60% lower than those kept in present day conditions, suggesting that conceptacle 42 development is initially highly sensitive to ocean acidification. However, this negative effect of ocean 43 acidification on conceptacle abundance disappears after three generations of exposure. Moreover, we 44 show that this transgenerational acclimation of conceptacle development is not facilitated by a trade-45 off with reduced investment in growth, as higher conceptacle abundances are associated with crusts 46 with faster growth rates. These results indicate that the potential reproductive output of coralline algae 47 may be sustained under future ocean acidification.

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49 Introduction

50 Ocean acidification is a threat to many marine organisms, with calcifying taxa and the ecosystems 51 they support particularly at risk [1,2]. As a hotspot for marine biodiversity [3,4], and an annual source 52 of billions of dollars to the global economy [5] coral reefs are one of the most valuable ecosystems 53 threatened by ocean acidification. These ecosystems are vulnerable because framework-building 54 corals and the coralline algae that cement this framework together may both be impaired by changes 55 in seawater carbonate chemistry [1,2]. Furthermore, coralline algae are predicted to be highly

56 susceptible to these altered conditions, as they precipitate one of the most soluble calcium carbonate 57 polymorphs, high magnesium calcite [6,7]. Coralline algae are important to the functioning of coral 58 reefs as they maintain the structural integrity of reefs by cementing them together, whilst also 59 providing a number of ecosystem services such as facilitating coral larval settlement [8,9] and 60 providing food for invertebrates [10]. Therefore, if coralline algal survival, growth or reproductive 61 output is reduced by future ocean acidification these services that are vital to the persistence of coral 62 reefs could be lost. Previous research has demonstrated reduced calcification and growth in mature 63 coralline algae under ocean acidification [11-13], as well as significant reductions in coralline algae at 64 natural analogues of future ocean acidification [14,15]. Moreover, coralline algal recruitment and 65 calcification during early life-stages appear to be highly susceptible to reduced seawater pH lower 66 than 7.91 as demonstrated through decreasing calcification rates and ~90% reduced recruitment rate 67 (or success) [16–18]. If such large reductions in recruitment are realized under ocean acidification, 68 coralline algal populations will inevitably experience significant declines, with severe negative 69 consequences for coral reefs [19].

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71 When assessing the impacts of environmental change on marine organisms it is important to consider 72 time scales that match the timespan organisms will have to respond to external changes, as previous 73 studies have demonstrated differing responses under long and short term exposure [20,21]. Such 74 considerations are also required as a host of transgenerational processes can act over multiple 75 generations to improve fitness. For example, tropical reef fish are initially sensitive to ocean warming 76 but can rapidly acclimate over two generations [22], whilst Rock Oyster larvae exposed to elevated 77 pCO_2 grew larger and developed faster when their parents also experienced elevated pCO_2 [23]. When 78 discussing generational improvements in performance it is important to define the use of terms such 79 as acclimation or adaptation. Here, we use the term acclimation to refer to changes in performance 80 driven by non-genetic plasticity in responses [22,24,25], and the term adaptation to refer to changes in 81 performance driven by genetic change [26,27]. A recent study by Cornwall et al., [18] was the first to 82 assess the response of a reef-accreting calcifier to ocean acidification across multiple (>two) 83 generations. They report that although coralline algae are initially highly sensitive to ocean

84 acidification, growth was no longer affected by ocean acidification after six generations of exposure. 85 Yet, this novel result should be interpreted with caution, as increases in growth after multiple 86 generations under ocean acidification may be at the expense of other processes such as reproduction. 87 Similar trade-offs have been observed in coralline algae, where maintaining calcification rates under 88 ocean acidification was coupled with harmful changes in skeletal ultrastructure [28,29], although such 89 trade-offs are not always observed [30]. Therefore, whilst measuring growth is valuable, a multi-90 generational assessment of reproductive structures in response to ocean acidification will provide 91 further insights into how coralline algae respond to ocean acidification.

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93 The pH variability within any one habitat can dictate the sensitivity of resident calcifying species to 94 ocean acidification. Exposure to variable pH regimes is common in shallow-water reef ecosystems 95 where local metabolic activity and long water residence times combine to cause large diurnal oscillations in seawater chemistry [31,32]. Such variable pH regimes may mitigate the effects of 96 97 ocean acidification on coralline algae by facilitating enhanced calcification rates during the day 98 [13,33], by selecting for individuals with low night pH tolerance due to phenotypic plasticity [34,35] 99 or natural transgenerational acclimation [25,36]. However, such hypothesised benefits are not always 100 apparent, with roughly half of past studies indicating coralline algae respond negatively to greater pH 101 variability [37], and that pre-exposure to low pH due to pH variability does not facilitate acclimation 102 to ocean acidification [18,38–40]. Contrasting results such as these suggest the effect of pH variability 103 on coralline algal responses to ocean acidification may be species or population specific. To-date, the 104 impacts of pH variability on the conceptacles of coralline algae from sites with vastly different pH 105 regimes has not been assessed

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107 To understand the vulnerability of coralline algal recruitment and early life stages to ocean

108 acidification, it would be ideal to isolate and investigate the various stages of reproduction and early

- 109 life stage development. Coralline algae theoretically have a triphasic life cycle that allows
- 110 reproduction via both sexual and asexual pathways. In coralline algae, like in many rhodophytes, the

111 asexual phase may be the only one ever observed [41] and occurs through tetraspore-to-tetraspore-112 cycling, in which tetrasporophytes produce self-perpetuating diploid spores [42]. Additionally, 113 tetrasporophytes have been observed to produce diploid bispores, which can also grow directly into 114 tetrasporophytes [42]. In coralline algae, many reproductive structures, including the spore producing 115 sporangia, are housed within reproductive conceptacles found on the surface of intergenicula or crusts 116 [42]. Thus, the development and abundance of these conceptacles (Figure 1) have a key role in 117 sustaining coralline algal populations and can be used as a proxy for reproductive output. However, it 118 is important to note that measurements of conceptacle abundance do not exactly equal fecundity. 119 Despite their importance, there has been limited research into the direct effects of ocean acidification 120 on coralline algal conceptacles, with the single study conducted finding no impact of reduced pH 121 (7.81, 7.60) on *Pneophyllum* sp. conceptacle numbers [43].

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123 Here, we investigate how reduced mean pH, differing levels of pH variability and the pH regime 124 experienced by coralline algae in their natural habitat impact the abundance and size of conceptacles 125 of the reef building coralline algae Hydrolithon reinboldii, across six generations of exposure. We 126 hypothesize that, (1) coralline algal conceptacle abundance and size will decline under ocean 127 acidification, (2) prior exposure to low pH will not influence H. reinboldii's tolerance to ocean 128 acidification, (3) diurnal pH variability will not reduce the effect of ocean acidification, and (4) any 129 impact of ocean acidification on H. reinboldii conceptacles will be reduced after multiple generations 130 of exposure.

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132 <u>Materials and Methods</u>

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134 Treatment Design and Experimental Set-up

135 Coralline algal rhodoliths of the species *Hydrolithon reinboldii* were collected from two sites with

136 contrasting pH regimes in the macrotidal Kimberley region of Australia in October 2016 (see figure

137 S1 for a map of these sites and figure S2 for information on the pH regimes). See supplementary

138 methods for site description and collection information. To investigate the effects of mean pH and 139 different levels of pH variability on *H. reinboldii* four pH treatments were employed within aquaria. 140 Two treatments had a mean pH_T of 8.00, selected to represent the present day mean surface seawater 141 pH_T detailed in ocean acidification studies best practices guide [44], and two treatments had a mean 142 pH_T of 7.70, designed to simulate potential future pH under ocean acidification (Representative 143 Concentration Pathway 8.5; 37). For the two treatments with the same mean pH, one was 144 characterised by high pH variability and the other low pH variability, designed to emulate the pH 145 regimes of Tallon and Shell Island respectively. Thus, the four experimental pH treatments were 146 denoted: (1) present-day low variability (8.00 mean pH_T, mean daily range of 0.14), (2) present-day 147 high variability (8.00 mean pH_T, mean daily range of 0.58), (3) ocean acidification low variability 148 $(7.70 \text{ mean pH}_{T}, \text{ mean daily range of } 0.20)$ and (4) ocean acidification high variability (7.70 mean 149 pH_{T} , mean daily range of 0.89). Figure S3 highlights the pH regime of all four treatments and Table 150 S1 displays the achieved pH for each of the four pH treatments used. Coralline algae collected from 151 the two sites were divided amongst the four treatments in 48 experimental tanks, resulting in six 152 independent experimental tanks for each of the eight pH treatment × site of origin combinations. See 153 supplementary methods for full details of experimental set-up and design.

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155 Multi-generational Experiment

156 In this study only the conceptacles of coralline algae that recruited during the experiment were 157 assessed, therefore the generation one mature rhodoliths collected from the wild (F0), were used 158 solely as source populations. The juvenile coralline algae assessed in this study are the same as those 159 studied in Cornwall et al., [18]. Whilst in the experimental tanks the generation one wild coralline 160 algae released spores which went on to form recruits on the inside walls of the tanks. After 78 days 161 these generation two coralline algal recruits were visible on the walls of the experimental tank. 162 Mature wild individuals were removed from the tanks on day 128 and two 4.5 cm \times 5.0 cm (length \times 163 width) sections were cut out of each tank. These two sections of each tank were primarily cut from the 164 area near the water outflow where possible, as this area contained the highest cover of coralline algal

165 recruits. These plates were then placed into completely new tanks (in the same treatment as the 166 previous generation), allowing for assessment of this second generation. Within these new 167 experimental tanks generation two algae formed conceptacles and released spores, which again settled 168 onto the walls of the tank. Two 4.5 cm \times 5.0 cm sections containing generation 3 recruits were cut 169 from the experimental tank and placed into a new experimental tank (See figure S4 for schematic representation of this experimental approach). This process was repeated every 41-51 days, allowing 170 171 five generations (generation numbers 2–6) of plate-based coralline algal conceptacles to be studied. 172 Coralline algal recruits across all generations possessed uniporate conceptacles, containing multiple 173 tetrasporangial spores which are characteristic of *H. reinboldii* [41], and suggests each new generation 174 was a result of asexual reproduction [46]. Contamination by foreign spores from local environments 175 was unlikely due to the consistent times to visible recruitment across generations, the absence of 176 coralline algae recruiting in the header tanks, and the observation that all recruits had similar 177 anatomical features [41]. To investigate whether any observed differences across generation were due 178 to an altered response to treatments or an interaction between altered response to treatments and 179 laboratory effects [47], a reciprocal transplant experiment was conducted. For full details of this 180 reciprocal transplant experiment see supplementary methods.

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182 Conceptacle Measurements

At the end of each generation, the coralline algae plates were removed from the tanks and stored in silica gel for up to three years, before three random locations on each plate were blindly selected and photographed using a compound microscope for analysis. For each plate conceptacle abundance was measured by counting all the visible conceptacles in the area of analysis. Individual conceptacle diameters were measured as the distance between the two most distant points of each conceptacle. For full details of microscopic conceptacle imaging and analysis see supplementary methods.



Figure 1. Images of the *Hydrolithon reinboldii* conceptacles quantified in this study. Shown are examples of the microscopic images used to quantify conceptacle abundance and diameter.

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199 Growth and Recruitment Measurements

200 Multi-generational growth and total recruit area of the same coralline algae populations were

201 investigated in relation to the conceptacle abundances measured in this study. Full methods and data

are reported in Cornwall et al., [18], but see supplementary methods for details of these measurements

and their comparison with conceptacle abundances.

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205 Statistical Analysis

206 Conceptacle abundance per plate and mean conceptacle diameter per plate was analysed using 207 generalised linear mixed models with generation, mean treatment pH, pH variability and site of origin 208 as fixed factors and header tank and water bath as random effects. As data was not normally 209 distributed a Poisson distribution was specified for conceptacle abundances and a Gamma distribution 210 was specified for conceptacle diameter measurements. For the reciprocal transplant experiment, both 211 conceptacle abundance and diameter were assessed using the same model structure, except that 212 generation and mean treatment pH factors were replaced with the factors; generation 2–6 mean 213 treatment pH and transplant treatment mean pH. Following the reciprocal transplant experiment, 214 Wilcoxon rank sum tests were used to test for pairwise differences in conceptacle abundances 215 between the different generation 2-6 and transplant treatment combinations.

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217	To allow for comparison with growth and recruitment data, mean conceptacle abundances per plate
218	were averaged for each experiment tank (total of 48 per generation). The effect of conceptacle
219	abundance on the total recruit area of the following generation, as well as the influence of treatment
220	mean pH was assessed using generalised linear mixed models with conceptacle abundance and mean
221	treatment pH as fixed factors. A quasi-Poisson model was used to account for non-normality of data
222	and overdispersion greater than 1. The same model was used to investigate whether changes in growth
223	rates impacted the abundance of conceptacles, however in this model growth and mean pH were
224	specified as fixed factors. Assumptions of all models were checked via visual inspection of standard
225	model diagnostic plots. All analyses were performed in R version 3.6.1.
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240 Results



Figure 2. *H. reinboldii* conceptacle abundance after exposure to present-day (mean pH 8.00) and ocean acidification (mean pH 7.70) treatments for between two and six generations. Figures are split by site of origin (Shell or Tallon Island) and variability of pH treatment (low or high). Median, 25% and 75% quartiles are presented.





in generations 2–6 (F = 7.797, p = 0.012) (Figure 3). However, transplant treatment (F = 0.059, p = 0.012) (Figure 3).

295	0.810) and the interaction between generation 2-6 treatment and transplant treatment (F = 0.303, p =
296	0.589) did not affect conceptacle abundance. Populations exposed to ocean acidification treatments
297	for seven generations had a mean conceptacle abundance of 4.04 conceptacles mm ² , whereas those in
298	present day treatments throughout generations 2-6, had a mean conceptacle abundance of 3.06
299	conceptacles mm ² after being transplanted to ocean acidification treatments in generation seven. This
300	represents a 25% decline in conceptacle abundance in populations transplanted to ocean acidification
301	treatments compared with those in ocean acidification treatments for seven generations. This
302	difference is demonstrated by a significant pairwise difference between conceptacle abundances in
303	these two reciprocal transplant groups (W= 38.5 , p = 0.024). Mean conceptacle abundance, standard
304	error and n for each generation 2-6 treatment by transplant treatment combination is detailed in Table
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323 Conceptacle Abundance vs Growth and Recruitment

324 In each generation an increase in mean coralline algal growth per tank has a positive effect on the 325 mean conceptacle abundance of the same coralline algae (Figure S6). Increasing conceptacle 326 abundance with increasing growth rates are evidence of this (F = 110.937 p < 0.001). The relationship 327 between growth and conceptacle abundance is not affected by mean treatment pH (F = 0.006, p =328 (0.937) or generation (F = 1.958, p = 0.163). As the conceptacle abundance of the plate-based coralline 329 algae increases, the total recruit area (of the next generation) derived from that population also 330 increases (F = 39.671, p < 0.001) (Figure S7). Mean treatment pH (F = 18.848, p < 0.001) and 331 generation (F = 7.610, p = 0.006) have a significant effect on total recruit area (F = 18.848, $p < 10^{-10}$ 332 0.001). The relationship between conceptacle abundance and next generation recruitment is not 333 affected by mean treatment pH (F = 0.006, p = 0.937) yet it is affected by generation (F = 5.100, p = 334 0.0250).

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336 (b) Conceptacle Size

Conceptacle diameter increased with generation (F = 14.099, p < 0.001, df = 1) (Figure S8), however
diameter did not vary with mean treatment pH, site of origin or level of pH variability (Table S7,
Figure S9). The reciprocal transplant experiment showed no effect of generation 2–6 treatment,
transplant treatment or any interaction between the two on diameter (Figure S10).

341

342 **Discussion**

343 Here, we demonstrate that coralline algal conceptacle abundance is initially impacted by ocean

344 acidification, but conceptacle size is not. This result indicates that the initial formation of

345 conceptacles, rather than their later development is impaired by altered carbonate chemistry.

346 However, as the first study to assess the impact of ocean acidification on conceptacles across multiple

347 generations, this study also provides evidence that coralline algae may be more robust to the effects of

348 ocean acidification than widely predicted, as from the third generation onwards there is no impact of

349 ocean acidification on conceptacles (Figure 2). Such an observation indicates that coralline algal

350 conceptacles will continue to develop in a more acidic ocean. This is important as conceptacles are

351 vital reproductive structures, without which the production of spores and subsequent production of 352 next-generation recruits would not be possible. Thus, if conceptacles were impacted by future ocean 353 acidification it would have severe consequences for the persistence and proliferation of coralline algal 354 populations. This ability to maintain conceptacle abundances and size under ocean acidification will 355 also be beneficial to other coral reefs organisms that are threatened by environmental change [1,2]. 356 For example, continuing spore production and the formation of new coralline algal crusts will provide 357 substrate for future coral larval settlement [8,9] and will also aid in binding calcium carbonate reef 358 structures together, as they are increasingly weakened by ocean acidification [7,28]. Therefore, 359 through such downstream effects the maintenance of coralline algal conceptacles under ocean 360 acidification could be vital to the future persistence of coral reefs.

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362 We consider it likely that the reduced effect of ocean acidification across generations observed in this 363 study is due to non-genetic acclimation, as coralline algae across all generations possessed features 364 that are indicative of asexual reproduction [46]. However, we cannot definitively state that the 365 changes observed here were not driven by selective genetic adaptation across generations. The 366 transgenerational acclimation of coralline algal conceptacle abundances to ocean acidification observed here follows the trend observed for growth rates [18]. Moreover, the positive relationship 367 368 between conceptacle abundances and growth rates (Figure S6) across all generations and treatments 369 demonstrates that trade-offs between these two processes are not underlying the multi-generational 370 acclimation of conceptacle abundance and growth to ocean acidification. For example, if such a trade-371 off was driving this acclimation you would expect to see a negative relationship between growth and 372 conceptacle abundance within the later "acclimated" generations. This observation that multi-373 generational acclimation is observed for conceptacle abundance and growth [18] further indicates that 374 coralline algae are capable of acclimating to ocean acidification. However, despite the overall 375 similarity in responses, acclimation of growth rates to ocean acidification were stepwise/linear in 376 nature (reported in Cornwall et al., [18]), and therefore not as rapid as the acclimation of 377 conceptacles. Whilst the acclimation of conceptacle formation observed here is deemed rapid 378 compared with growth, similar rapid transgenerational acclimation has been observed in differing

379 marine taxa. For example, the reproductive output of marine invertebrates such as polychaetes [48] 380 and copepods [49] is initially impaired under ocean acidification, however this effect disappears by 381 the third generation, as the reproductive output of populations exposed to ocean acidification does not 382 differ from controls. Additionally, transgenerational acclimation of growth rates to ocean acidification 383 has been observed within three generations of exposure in calcifying organisms such as oysters [50] 384 and clams [24], suggesting various biological processes are capable of rapid acclimation to ocean 385 acidification. However, our results and those of Cornwall et al., [18] suggest that different processes 386 may exhibit contrasting acclimation times.

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388 The varying rates of acclimation of different biological processes in coralline algae are likely driven 389 by the physiological mechanisms underpinning these processes. Here, we propose two mechanisms 390 that may have driven the faster multi-generational acclimation of conceptacle development, compared 391 with growth [18]. Firstly, linear extension of coralline algae is dictated by both growth of organic 392 tissue and calcification of the cell walls. The latter is impacted by factors such as the control of 393 calcification site chemistry and the production of organic matrices, both of which may be disrupted by 394 ocean acidification [51,52]. Thus, the complexity of calcification and its dependence on specific 395 chemical conditions may represent a physiochemical bottleneck to the rate at which calcification, and 396 therefore growth can acclimatise to ocean acidification. In contrast, the primarily uncalcified [53,54], 397 conceptacles are not subject to the same chemical constraints, potentially facilitating faster 398 acclimation. An alternate and not mutually exclusive explanation is that coralline algae prioritise 399 energetic investment in reproduction over calcification. Energetic trade-offs between reproduction and 400 other processes such as growth commonly occur in nature [55,56], and are enhanced due to increased 401 maintenance costs under stressful conditions [57]. Therefore, rapid acclimation of reproductive 402 conceptacles may come from prioritizing energetic investment in reproduction, as ensuring the 403 production of offspring under stress is of the greatest benefit to Darwinian fitness. Similar multi-404 generational trends in resource allocation towards reproduction under ocean acidification have also 405 been observed in copepods [58]. Here, our novel comparison of acclimatory responses (growth rates

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409 The reduction in second generation conceptacle abundances under ocean acidification mirror the 410 widely reported negative impacts of ocean acidification on the calcification, growth and recruitment 411 of first and second generation coralline algae [11,16,59]. This initial reduction in conceptacle 412 abundance observed here, highlights reduced conceptacle formation as one potential mechanism 413 driving these previously reported declines in coralline algae recruitment under ocean acidification 414 [16,17]. However, in contrast to the initial negative effect observed here, Ordoñez et al., [43] found 415 that conceptacle numbers in *Pneophyllum* sp. coralline algae were not reduced when exposed to 416 medium (pH 7.81) or low (pH 7.60) pH treatments. Such a result is unexpected, as these populations 417 were subject to neither transgenerational nor developmental exposure to low pH conditions, as they 418 recruited into experimental conditions from the wild. However, species-specific responses to ocean 419 acidification have been documented in coralline algae [37,60], therefore this opposing result may be 420 due to differences between the thin crusted *Pneophyllum sp.* studied by Ordoñez *et al.*, [43], and the 421 thick-crusted H. reinboldii assessed here. Evidence of such species effects are displayed in Ordoñez et 422 al., [43] as they observed a significant decline in the relative density of thick crusted Porolithon 423 onkodes under high pCO₂, but not *Pneophyllum sp.*, suggesting that morphology can influence 424 response to ocean acidification. Alternatively, these populations may have had natural resilience to 425 this magnitude of ocean acidification, and further research is required to elucidate which factors could 426 be driving such differences.

vs conceptacle development), indicates that the persistence of populations under ocean acidification

will be constrained by the biological process that undergoes the slowest rate of acclimation.

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The rapid acclimation of reproductive conceptacles observed here is not translated to rapid increases in final recruit area of the next generation, as recruit area acclimation was slower and stepwise [18]. Therefore, although the trend of increasing recruit area with increasing conceptacle abundance (Figure S7) suggests that higher conceptacle abundances may aid in buffering the impacts of ocean acidification on recruitment, it is likely that final recruitment will be determined by the cumulative impact of ocean acidification on reproductive structure formation, spore settlement, and early

434 development processes. For example, coralline algae display delayed and weakened spore attachment

435 [61], as well as reduced spore germination [62] in response to ocean acidification, potentially

436 highlighting the later sporulate stage as a bottleneck in final recruitment.

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438 Our results also demonstrate that pH variability does not modify the impact of ocean acidification on 439 coralline algal conceptacles. This is shown as coralline algae grown in both high variability and low 440 variability pH treatments were not differently affected by the mean pH corresponding to ocean 441 acidification. This lack of effect of pH variability is at odds with some past literature, where pH 442 variability has been shown to reduce the impact of ocean acidification [33,38], while it can also 443 enhance the negative impacts of ocean acidification [37]. These contrasting and inconsistent effects of 444 pH variability demonstrate that its interaction with future ocean acidification will be complex and can 445 potentially elicit a full range of effects from negative to positive, with the lack of effect observed here 446 in between the gamut of responses. As previous studies have highlighted that the influence of pH 447 variability varies amongst species [20,63], it appears that the effect of ocean acidification on 448 *H.reinboldii* is not dictated by pH variability as the calcification and photosynthesis of mature 449 rhodoliths [40], the growth rates and recruitment of juveniles [18] and the size and abundance of 450 juvenile conceptacles are all unaffected.

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452 The effect of ocean acidification on coralline algal conceptacles is not influenced by the pH variability 453 experienced by the population in their natural environment here. This was hypothesized a priori, as 454 similar responses were observed with adult calcification rates [40] and juvenile growth rate [18]. 455 Nevertheless, this result contrasts with the often-posed hypothesis in the literature that environments 456 with naturally high variations in an abiotic factor, will increase the resistance of resident organisms to 457 future mean changes in that abiotic factor [37]. Though this trend has recently been observed in corals 458 from habitats with greater thermal variability [64,65], mean conditions may be as important or more 459 important than variability [66]. For example, mean pH is higher at Tallon Island, while pH variability 460 is greater than Shell Island. Therefore, both could be acting against each other in conditioning 461 coralline algal responses to low pH. Our results suggest a similar priming effect of mean pH, as three

generations of exposure to reduced mean pH was enough to eliminate the effects of reduced mean pHon the conceptacles of later generations.

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465 Overall, this study demonstrates that the abundance of coralline algal conceptacles is greatly reduced 466 in second generation coralline algae reared under simulated ocean acidification conditions. However, 467 in just three generations of exposure to these same ocean acidification conditions the coralline algal 468 reproductive structures can rapidly acclimate, as we observe no effects of ocean acidification beyond 469 the second generation. This acclimation is much faster than that previously reported for growth rates, 470 as growth took six generations to acclimate. Despite differing rates of acclimation, the acclimation of 471 conceptacle development observed here is not facilitated by reduced growth, as both processes 472 acclimate together across all generations. Therefore, this study further supports the possibility that 473 coralline algae could acclimate to the effects of ocean acidification. As multiple studies have reported 474 the ability of marine organisms to acclimate to environmental change, our observation that 475 acclimation rates vary for different biological processes introduces a new consideration for studies 476 assessing how marine organisms will respond to future environmental change. 477 478 **Ethics** 479 All local regulations and permit requirements were followed during this study. 480 481 **Data accessibility**

- 482 Data is accessible in the Dryad Digital Repository:
- 483 https://datadryad.org/stash/share/BVDT9xgRvmfMvs-ruppJEYdEBSB81PoKG8c_1UmetTQ

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485 <u>Authors Contributions</u>

- 486 B.M. wrote the paper. C.E.C and S.C. conceived the research and designed the experiment. C.E.C.,
- 487 S.C. and B.M. conducted fieldwork. B.M., C.E.C., S.C., F.P. and E.L. ran the experiment. B.M., A.C.,
- 488 M.B. and A.P. conducted photograph analysis. M.T.M. provided essential laboratory equipment,

489	facilities and analyses. B.M. conducted the statistical analyses. All authors edited the manuscript, or
490	provided intellectual input, and agreed to its submission.

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499 Conflict of Interest

- 500 The authors declare no conflicts of interest.
- 501

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